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# Life cycle assessment of biofuels produced by the new integrated hydropyrolysis-hydroconversion (IH $^2$ ) process

Edwin Maleche Michigan Technological University

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### LIFE CYCLE ASSESSMENT OF BIOFUELS PRODUCED BY THE NEW INTEGRATED HYDROPYROLYSIS-HYDROCONVERSION (IH<sup>2</sup>) PROCESS

By

### EDWIN MALECHE

A THESIS Submitted in partial fulfillment of the requirements for the degree of

> MASTER OF SCIENCE (Chemical Engineering)

### MICHIGAN TECHNOLOGICAL UNIVERSITY 2012

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This thesis, "Life Cycle Assessment of Biofuels Produced by the New Integrated Hydropyrolysis-Hydroconversion  $(IH^2)$  Process," is hereby approved in partial fulfillment of the degree of MASTER OF SCIENCE IN CHEMICAL ENGINEERING.

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# **Table of Contents**







# **List of Figures**







**Figure 6.12:** Network diagram for forest resources  $IH^2$  biofuels emission of GHGs (kg  $CO<sub>2</sub>$  eq./MJ IH<sup>2</sup> biofuels). Red lines show relative magnitude of greenhouse gas impacts while green lines show credits due to co-products of production. Line width corresponds to magnitude of impact or credit. ..50



Figure 7.10: The process flow diagram for micro algae production from Aquaflow Bionomic Corporation. The shaded boxes represent the main stages of production………………………………………………………………………..71

# **List of Tables**



**Table 6.5:** Estimated distances for different blending locations……………….…....32



- **Table 6.14**: Greenhouse Gas Emissions per dry metric ton/day of wood and forest residues collected, transported, and processed on-site. Impacts of all greenhouse gases were converted to  $CO<sub>2</sub>$  equivalents using Global Warming Potentials (GWP). Plant sizes of 500 and 1000 dry metric ton/day input feedstock considering electrical energy from US average grid as the yard processing energy source and assumption of \$3 per gallon of diesel fuel used. .................................47
- **Table 6.15**: Greenhouse Gas Emissions per dry metric ton/day of wood and forest residues collected, transported, and processed on-site. Impacts of all greenhouse gases were converted to  $CO<sub>2</sub>$  equivalents using Global Warming Potentials (GWP). Plant size of 500 and 1000 dry metric ton/day input feedstock considering electrical energy from US average grid as the yard processing energy source and assumption of \$6 per gallon of diesel fuel used……………… ..........48
- **Table 6.16:** GHG emissions for the  $IH^2$  process with 50% moisture feedstock content. .50
- **Table 6.17:** GHG emissions for the  $IH^2$  process with 30% moisture forest resources assuming 100 km transport of IH2 biofuel by different modes ..............................51
- **Table 6.18**: Estimated distances for different blending locations .....................................51



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# **List of Symbols/Abbreviations**



# **Abstract**

Biofuels are alternative fuels that have the promise of reducing reliance on imported fossil fuels and decreasing emission of greenhouse gases from energy consumption. This thesis analyses the environmental impacts focusing on the greenhouse gas (GHG) emissions associated with the production and delivery of biofuel using the new Integrated Hydropyrolysis and Hydroconversion  $(IH^2)$  process. The IH<sup>2</sup> process is an innovative process for the conversion of woody biomass into hydrocarbon liquid transportation fuels in the range of gasoline and diesel.

A cradle-to-grave life cycle assessment (LCA) was used to calculate the greenhouse gas emissions associated with diverse feedstocks production systems and delivery to the  $IH<sup>2</sup>$ facility plus producing and using these new renewable liquid fuels. The biomass feedstocks analyzed include algae (microalgae), bagasse from a sugar cane-producing locations such as Brazil or extreme southern US, corn stover from Midwest US locations, and forest feedstocks from a northern Wisconsin location.

The life cycle greenhouse gas (GHG) emissions savings of 58%–98% were calculated for IH<sup>2</sup> gasoline and diesel production and combustion use in vehicles compared to fossil fuels. The range of savings is due to different biomass feedstocks and transportation modes and distances. Different scenarios were conducted to understand the uncertainties in certain input data to the LCA model, particularly in the feedstock production section, the IH<sup>2</sup> biofuel production section, and transportation sections.

# **Keywords**

- Life cycle assessment
- $\cdot$  IH<sup>2,</sup> Integrated Hydropyrolysis and Hydroconversion process;
- Biomass
- Greenhouse gas emission
- System Boundary
- Intergovernmental Panel on Climate Change

# **1. Introduction**

Gas Technology Institute (GTI) has developed an innovative process for the conversion of woody biomass into hydrocarbon liquid transportation fuels in the range of gasoline and diesel. The process for this conversion is referred to as "Integrated Hydropyrolysis and Hydroconversion,  $IH^{2\nu}$ . The environmental impacts of producing and using these new renewable liquid fuels are largely unknown, and therefore, MTU was contracted to conduct a cradle-to-grave Life Cycle Assessment (LCA) of these new biofuel products. In addition, several biomass feedstocks were included in the scope of the requested LCA, because it is anticipated that the  $IH^2$  will be able to accommodate a variety of biomass feedstocks. The biomass feedstocks include algae (microalgae), sugar cane bagasse, corn stover, and forest feedstocks from a location in the Upper Midwest (Wisconsin). This report contains a preliminary LCA of  $IH^2$  biofuels based on input data for the production and delivery of biomass feedstocks to a future biofuel facility, and also based on inputs for the  $IH^2$  process provided by GTI.

# **1.1 Background on the IH<sup>2</sup> Process**

A process flow diagram of the  $IH^2$  process is shown in Figure 1.1. A detailed description of the IH<sup>2</sup> process can be found in GTI publications such as in (Marker et al. 2009). The process is carried out in two sequential yet integrated stages at moderate pressure (250- 500 psi); hydropyrolysis and hydroconversion. Briefly, the process is carried out in two integrated steps: hydropyrolysis and hydroconversion. The first step is an exothermic catalytic fast hydropyrolysis reaction carried out in a fluid bed reactor at moderate hydrogen pressure. The product vapors from the first step are carried to the second conversion step, a hydrodeoxygenation reactor operating at essentially the same pressure as the first hydropyrolysis reactor. The hydrogen required for the  $IH^2$  process is produced in a reformer using C1-C3 co-products, and therefore no external hydrogen source is needed, such as  $H^2$  from steam reforming of methane. Other by-products of the process



**Figure 1.1 Process flow diagram of IH<sup>2</sup> process (Marker et al. 2009)** 

are char, high pressure steam, and ammonia / ammonium sulfate (not shown in Figure 1.1). Ammonia and hydrogen sulfide in the process condensate are stripped and oxidized to make ammonium sulfate which can be used as a fertilizer. More detailed diagram is shown in the appendix G.

### **1.2 Background on Feedstocks**

Biomass types for this LCA were diverse representing feedstocks from forest, agricultural, and aquatic environments. These biomass types include algae (microalgae), bagasse from a sugar cane-producing location such as Brazil or extreme southern US, corn stover from a Midwest US location, forest feedstocks from a northern Wisconsin location. Inputs for the production, preparation, delivery, and storage of these biomass feedstocks were provided by several industrial partners in this project, as discussed later in this report. From this input data, we conducted a LCA of just the biomass production system from the "field" to the input of the  $IH^2$  process. These analyses were useful to not only compare and contrast different feedstocks for biofuel production, but also to recommend steps to reduce the environmental impacts of such feedstock production systems.

### **1.3 LCA Research Objectives**

The main research objectives for this report are;

- 1. Conduct a cradle-to-gate LCA of different biomass feedstocks for  $IH<sup>2</sup>$  biofuel production.
- 2. Conduct a cradle-to-grave LCA of  $IH^2$  biofuels produced from different biomass feedstocks.
- 3. Investigate uncertainties in LCA inputs through scenario analyses.

The following sections of this report will provide details on the LCA methods used, on the input data included in the analysis, and on the greenhouse gas emissions of  $IH<sup>2</sup>$ biofuels. Comparisons will be made to petroleum fuels with respect to savings of GHG emissions over the  $IH^2$  biofuel life cycle.

### **2. LCA Methods**

LCA is mainly used to determine the environmental effects and performance of a product over its full life cycle. Motivation for using LCA in this current MS thesis is to not only satisfy the demands of the research sponsor (GTI), but also to enumerate the greenhouse gas emissions according to methodology from regulatory agencies such as the Environmental Protection Agency. LCA identifies the emissions and energy savings and aids in research and development. The LCA approach is chosen for this project, because it easily avoids having a narrow environmental concern outlook.

Alternative bio-based transportation fuels have the potential to decrease climate change emissions from vehicular transportation. The magnitude of this emission reduction can best be determined using the methods of life cycle assessment (LCA) by considering the entire life cycle of the new biofuel product from biomass cultivation through conversion to biofuel product, and use in vehicles. The methods for LCA put forth by the International Organization for Standardization (ISO 2006) were followed in this analysis. The main steps in conducting a life cycle assessment are as follow, and further details on each step will appear later in this report.

- Life cycle goal and scope and functional unit definition
- Life cycle inventory analysis
- Life cycle impact assessment
- Life cycle interpretation

### **2.1 Goal and Scope and Functional Unit**

The main purpose of this research project is to help GTI develop a better and more sustainable biofuel manufacturing process and product. This is done by estimating the environmental burdens in the form of GHG emissions that are associated with the GTI  $(H<sup>2</sup>)$  biofuel production process. In satisfying this main purpose, this study will evaluate the cradle-to-grave life cycle assessment (LCA) of the Gas Technology Institute (GTI) Integrated Hydropyrolysis and Hydroconversion  $(H<sup>2</sup>)$  production chain, including the production of input feedstocks and use of output  $I\!H^2$  biofuels. The goal is to compare environmental impacts of  $IH^2$  biofuels to equivalent fossil fuels in order to determine savings of emissions, but along this path, intermediate results for each biomass feedstock will be generated and compared to each other. The scope of this LCA will be from cradle-to-grave and the impacts of concern are greenhouse gas emissions. The functional unit for biomass feedstocks and  $IH^2$  biofuels will be 1 dry metric ton and 1 MJ of energy, respectively. The input data for these LCAs will be organized by the scale of production; 1 dry metric ton for biomass inputs, and for  $IH^2$  biofuels production, 2,000 moisture and ash-free (MAF) metric tons/day facility. The LCA results for  $IH^2$  biofuels were generated by dividing the LCA emissions by the total energy content in MJ of  $IH<sup>2</sup>$  biofuel produced each day from the facility. This biofuel production changed depending on the specific biomass input feedstock input to the facility, as shown in the subsequent sections.

### **2.2 Life Cycle Diagram and System Boundary**

The life cycle diagrams describing each  $IH^2$  biofuel production system is presented in Figure 2.1 for microalgae, sugar cane bagasse, corn stover, and forest feedstocks. Each diagram has similarities and subtle differences, especially in the feedstock production stage, the first stage on the left of each diagram. Fuels, electricity, lubricants, and grease are common inputs for each of the feedstock production stages because of the presence of machines for biomass collection and equipment for pumping algae solutions (microalgae) and for size reduction (bagasse, stover, and forest feedstocks). Fertilizers are required for stover feedstocks because this feedstock is produced from intensive agricultural practice which involved application of inorganic and organic fertilizers. When this feedstock is collected off of the land, the nutrients are removed with and must be replaced for successful subsequent crop production. At the  $IH^2$  conversion to biofuels stage, inputs of catalysts, electricity, and other chemicals are included, and outputs of co-products steam, ammonia, and ammonium sulfate are produced. Diesel fuel for transportation of  $IH^2$ biofuels to locations of blending into fossil fuel stocks is included, and consideration is given to transport to filling stations and also for emissions of greenhouse gases from vehicle use of the biofuels.

The next section will present tables of input data for production of biomass feedstocks and also of IH<sup>2</sup> biofuels produced from these feedstocks.

This analysis also considers the land use change which could both directly and indirectly affect the impact analysis. The direct land use change is where food crop land is converted to grow biofuel crops. During the land preparation step, some additional GHG emissions may occur and any carbon stock changes before and after establishment of bioenergy crops will alter the GHG analysis by contributing additional  $CO<sub>2</sub>$  sequestration or increasing GHG emissions. The indirect land use change is where agriculture land is converted to grow biofuel crops instead of food crops. Because food demand is "inelastic" meaning that food demand must be met as the highest priority, when agricultural lands in food production are diverted to bioenergy crop production, somewhere in the world natural wild lands will be converted to food production, with associated land use change GHG emissions.

Although land use change emissions were considered in this study, it was concluded that these effects are negligible for the following reasons. First for micro algae feedstock, this was acquired from natural ponds with runoff water from agricultural land practices, thus this had neutral effect to the land use change. Secondly for the corn stover, the feedstock was acquired from the farms as a waste within sustainable practices such as soil quality is not diminished, and therefore no extra land was required to satisfy food production, leading to having no effect to land use change. Thirdly, the bagasse feedstock was acquired as a waste from the sugar cane processing facility; therefore no land use change resulted from this use bagasse. Lastly, the forest resource biomass was considered as residue and therefore no extra land was required, leading to having no effect on land use change.









**Figure 2.1 Life cycle diagrams for production of IH<sup>2</sup> biofuels from different biomass feedstocks.** 

### **3. Life Cycle Inventory**

The life cycle inventory is the list of emissions associate with each input to the  $IH<sup>2</sup>$ biofuel life cycle. The total inventory is the sum of emissions for all of the inputs. The inventory of emissions resides within input-specific ecoprofiles in the ecoinvent database in SimaPro 7.2, the LCA software tool used in this study. For example, if diesel fuel is one input to the biomass feedstock production stage, an ecoprofile in the ecoinvent database in SimaPro 7.2 has a list of emissions inventory data for the production of this diesel fuel. We created a diesel combustion emission ecoprofile with an emission factor of 3.17 kg  $CO<sub>2</sub>$  / kg petroleum diesel combusted based on stoichiometry. Similarly, other ecoprofiles were used for other life cycle inputs such as transport by road (includes combustion emissions of diesel fuel), for fertilizer inputs, chemicals used, and catalysts. These inventories have data for calculation of many categories of environmental impact, but in this study the primary and sole category of interest is greenhouse gas emissions and global warming. The emissions inventory of the greenhouse gases  $CO_2$ ,  $N_2O$ ,  $CH_4$ , refrigerants, and solvents is therefore of primary interest. This study did not include the N2O emissions associated with nitrogen (N) fertilizers allocated to corn stover and cane bagasse production because the removal of N with these biomass feedstocks will have the effect of reducing  $N_2O$  emissions compared to the business-as-usual case (feedstocks left on the land to decompose and emit  $N_2O$ . This emissions reduction is compensated for when additional N fertilizer is applied to the subsequent corn and sugar cane crops in equal amounts. This assumption is justified based on "Tier 1" emission factors used in the Intergovernmental Panel on Climate Change (IPCC) (Eggleston et al. 2006).

### **3.1 Inputs for Biomass Feedstock Production**

#### **3.1.1 Inputs for Microalgae Production**

Table 3.1 below shows the algae production inputs used for the life cycle assessment for the Aquaflow Bionomic Corporation (ABC). This data was obtained from a spreadsheet provided by ABC based on Blenheim site Power assuming 100 g algae/m<sup>3</sup> cell density. The data was then divided into different sections. The first section was the raw material section which includes use of fertilizers which are all provided by the sewage plant or natural water body. The second section is the Pump Shed, which includes the supply and the discharge pumps; 5 electric motors whose energy use is measured in kWh/kg dry algae recovered. The third section is the New Harvest Unit. This section contributes much of the energy and is a total of 6 motors. The fourth section is the De-watering process section where several activities take place including removal of excess water by draining and rising which is done using electrical motors. The other important activity that takes place in this section, is use of chemical additives to agglomerate the algae at the dewatering stage to enhance the harvesting process. Lastly is the transportation to the IH<sup>2</sup> processing which is assumed to be done over a 100 km distance. The moisture in the algae was taken into account for this transport step assuming 80% moisture content.

The main inputs in Table 3.1 for the LCA analysis of the GHG emission was the electricity used by the motors at the pump shed section and new harvest unit section. Greenhouse gas emissions per kWh of electricity used were obtained from the US Environmental Protection Agency eGRID website assuming a U.S. average grid (US

#### **Table 3.1**



**Data inputs for algae cultivation, harvesting, and transport for Aquaflow Bionomic Corporation. Basis is 1 dry metric ton microalgae and 100 g algae / m3 .** 

EPA 2011) in the base case analysis. The emissions in this eGRID database are for electricity production only and do not include upstream process of production of primary energy (coal, etc.). To account for this, 10% extra emissions were added for these upstream processes. These additional emissions were arrived at after review of several electricity generation ecoprofiles in the ecoinvent database in SimaPro. The data provide by the ABC in Table 3.1 was divided by three so as to get the algae cell density of 300 g algae/m<sup>3</sup> because the original data was for 100 g algae/m<sup>3</sup> cell density. The process flow diagram of production of micro algae is shown in appendix I.

#### **3.1.2 Inputs for Bagasse Production**

Bagasse is considered a waste from the sugar or cane ethanol production process, and in this analysis it is assumed available with no environmental burden from its production. However, environmental impacts accumulate from bagasse handling in the  $\overline{IH}^2$  biofuels production life cycle. The step wise process of bagasse handling as a feedstock includes loading, transportation, and unloading to the  $IH^2$  facility. The first stage involves using a diesel powered front loader to transfer bagasse into trucks for transport. The second stage is the transportation stage, where the bagasse is transported using a 16-32 ton truck to be delivered to the IH<sup>2</sup> facility. The third stage is the unloading of the bagasse to IH<sup>2</sup> facility storage, and finally loading into the  $IH^2$  facility.

The main inputs in Table 3.2 are for loading/unloading and for transportation, which involves the use of 16-32 ton trucks to the  $IH^2$  facility. The bagasse may be ground to decrease the size so as to have the desirable size for the  $IH^2$  processing. The first stage is the loading of unbaled bagasse using front loaders directly from the bagasse piles at sugar milling factory onto trucks. There are three such loading/unloading steps and this is the cause of the factor of 3 in the inputs of Table 3.2 for diesel fuel. The factor of 1.1 converts from short tons, the basis for the input data from (Morey et al. 2010), to metric tons, and the factor of 1.45 accounts for the field moisture content of the bagasse, assumed to be 45%. The (Morey et al.2010) study was on corn stover, but the steps in the feedstock supply chain and equipment used are very similar to the bagasse supply chain, and therefore the use of this source of input data is justified. Drying of bagasse prior to entering IH<sup>2</sup> reactors is not included in this input data, but is included in the IH<sup>2</sup> process analysis section. There is not factor of 3 for lubricating oils because the input value includes this already. Emissions for combustion of diesel fuel is included in the analysis for loading / unloading steps using stoichiometric factor of 3.17 kg  $CO<sub>2</sub>$  / kg diesel combusted. Diesel volume in gallons was converted to kg by using a density of 0.85 kg diesel / L diesel and converting between gallons and liters.

#### **Table 3.2**

#### **Inventory data for bagasse loading, transportation, and unloading on a basis of 1 dry metric ton of feedstock.**



#### **3.1.3 Inputs for Corn Stover Production**

Corn stover feedstock production includes collection from the fields, loading, transportation, unloading, and fertilizer replacement to the fields to compensate for nutrients removed with the stover. We assume that there will be no change in soil organic carbon due to removal of some, but not all, of the stover from the field, and therefore no emissions of  $CO<sub>2</sub>$  from C stock change. The first stage in Figure 1 involves dieselpowered stalk shredder equipment used for shredding of the corn stover. Then the stover is collected, which involves raking and baling, and processed into round bales. Next is stover loading, where the round bales are lifted and moved using a front loader onto trucks for transportation. Then, the corn stover is transported and delivered to the  $IH<sup>2</sup>$ facility, and then finally unloaded to the storage area.

This analysis assumes 70% corn stover removal per unit land area with collection every other year that corn is grown, resulting in an average stover removal of 35% of area per year. This leads to more efficient, and less costly collection process and less soil compaction than harvesting of 35% of the corn stover each year. Lastly this analysis assumes that there is nutrient replacement to the corn stover harvested fields. Fertilizers rich in nitrogen, phosphate and potassium are used to replenish the nutrients lost from the field so as to have adequate nutrients for the growth of the next corn crop.

Table 3.3 shows LCA inputs for corn stover handling from the corn field to the  $IH<sup>2</sup>$ process as obtained from a recent research article (Morey et al.2010) and (Maleche et al. 2011). One of the key inputs is the nutrient replacement. The replacement fertilizers used are diammonium phosphate, ammonia solution, and potassium sulfate. The main diesel input in this process is during the stover collection stage, which involves stalk shredding, raking and baling. The stalk shredding occurs after harvesting of the corn and involves decreasing the size of the stalks by use of a mechanical shredder, which is diesel powered. The shredding is done so as to increase the volume of harvested corn stover and facilitate drying to the target moisture content of 15-20%. The shredded corn stover is then raked using a diesel powered machine. Lastly the stover is baled into round bales for easy handling and transport. The collection stage is the most critical step due to finding the suitable time period for the shredding, racking and round bailing of the corn stover with 15-20% moisture. The third main stage is the transportation stage, in this stage the stover in the form of round bales is loaded onto and transported by truck (25-ton). The last stage in this process is the unloading of stover bales to storage, and then loading of stored stover into the  $IH^2$  process. Transport distance by truck to the  $IH^2$  facility from the field is on average 30 miles (Morey et al. 2010).

#### **Table 3.3**



**Inventory data for the corn stover with a basis of 1 dry metric ton of feedstock. Each fuel and lubricant entry in this table is divided by 0.85 to convert to dry basis.** 

### **3.1.4 Inputs for Timber Resources Production**

Mr. John Gephardt has developed a model of timber resource procurement for northern Wisconsin on behalf of Johnson Timber Company (JTC) and provided information on the quantities of fuel, lubricants, and electricity based on the amount of feedstock delivered per day. This model was based on a wide range of available woody feedstock that were identified around a site located in Park Falls, Wisconsin. Types of feedstock included are: logging residues; un-merchantable timber; un-marketable timber; marketable timber; and mill residues. Each feedstock type has unique requirements in their collection, transport, and processing needs. Within any one type, quantities were available at differing distances to Park Falls. Based on the delivered costs for each feedstock the JTC model selects a blend of feedstock which would result in the lowest possible total costs for each plant size that was evaluated. The price of diesel fuel was included as a variable in the model. This allowed the model to take into account how the blend of feedstock in the output would be influenced as diesel prices change.

The stepwise process of wood and forest residue production in JohnsonTimber Company is illustrated in the flow sheet below in Figure 3.1. The first stage is the collection of resources from the forest. The processes involved in this stage include skidding and cutting of the biomass from the forest to the required length for transportation, roadside chipping and debarking, and loading of the round wood, slabs and chips using a log loader and chip dumps. The second stage is road transport in which the round wood, bark, sawdust, slabs, fuel rods, and woodchips are transported for processing to the  $IH^2$ facility. The last stage is the processing stage. In this stage size reduction occurs whereby there is conversion of the round wood and other sized biomass into chips small enough for the  $IH^2$  process. This stage also includes the use of grinders which can be either stationary (electrical powered) or mobile (diesel powered). In this analysis the grinders are assumed to be either stationary or mobile and are electric-powered according to information from Mr. Gephardt. In the last stage we have the mixing loaders which are used to blend the various types of feed stock which use screens to remove the oversized materials to the  $IH^2$  process.

The JTC model was used to evaluate biomass inputs rates ranging from 50 to1,750 dry short tons/day. Figure 3.2 shown below illustrates how the percentages of hardwoods and softwoods changed with increasing plant size. Within the supply area, hardwoods comprise approximately 70% and softwoods 30% of the available feedstock. The higher percentage of hardwood at the smaller plant sizes is the result of low valued hardwood residues available from an adjacent pulp and paper mill. For the study plant sizes of 500 dry short tons/day and 1,000 dry short tons/day of feedstock were selected for evaluation. The feedstock selected for each plant sized was values were chosen from an economic stand point. Figure 3.3 shows the distribution of total diesel fuel among feedstock collection, transportation, and processing (chipping). Above 1,000 dry short tons/day, there is not much change in total diesel consumption per dry short ton.

Table 3.4 and 3.5 show the wood and forest residue production inputs used for the life cycle assessment for the Johnson's Timber Company. This data was based on an assumption of \$3.00 and \$6.00 per gallon of diesel fuel in two separate scenarios. This data was divided into different sections. The first section involved the raw material collection which includes the use of lubricants, fuel, grease, hydraulic fluid, and gasoline. The second main section is the transportation which includes the use of lubricants and fuel. The third main section is the yard processing section. In this section several activities take place including wood chipping, screening, and conveying. These inputs include electricity for running the motors, and fuel and lubricants inputs for the different yard equipment.

The main data inputs in Table 3.4 and 3.5 are the diesel used for the collection and transportation of the wood to the  $IH^2$  processing plant. Lubricants and hydraulic oil values were assumed based upon the diesel consumption estimates provided by Mr. Gephardt on behalf of JTC. The fertilizer and other additives are assumed to be negligible because no use of these inputs occurs for timber cultivation. The main biomass feed stock inputs are underutilized round wood sources and the non-commercial tree species, since they are undesirable in the manufacturing of traditional forest products**.** Lastly the other main biomass feedstock inputs are forest residues which include tops, limbs and fuel rods. The fuel rods are defined as the round woods that do not meet the size and quality standards for traditional forest products and examples of this are the oversized and undersized stems from saleable and unsaleable trees.

In this inventory the second major input is the electricity used for the size reduction which is used in the electric motors of the stationary chipper. The materials which require high energy for size reduction are the sawmill slabs, fuel rods, and round woods which go through extensive processing for the size reduction. The main equipment used in the yard is the stationary chipper, conveyor system, over size screen, secondary hog and chip dumps. On the other hand, there are materials which do not require a lot of energy for size reduction due to be ready to use or being available in fairly small size particles.



**Figure 3.1 Process flow diagram for wood and forest residue production from Johnson Timber Company. The shaded boxes represent steps which are not included in the analysis presented here.** 



**Figure 3.2 The percentages of hard wood and soft wood used as the feed stock input with varying plant size.** 



**Figure 3.3 Diesel fuel consumption for collection, trucking, and processing as a function of biomass input rate.** 

#### **Table 3.4**

**Data inputs for wood and forest residue raw material collection, transportation and yard processing based on 1 dry short ton biomass with an assumption of \$3 per gallon of diesel fuel.** 



# **3.2 Inputs for IH<sup>2</sup> Biofuels Production**

# **3.2.1 Inputs for Microalgae IH2 Biofuels Production**

Table 3.6 shows the IH<sup>2</sup> facility inputs and outputs provided for the life cycle assessment. The data was obtained from Terry Marker (GTI) and was based on a 2,000 dry metric ton/day plant. The accuracy of input data was verified by carrying out a mass and energy balance as shown in appendix A. This data was based on an assumption of 20% moisture content of the microalgae biomass feedstock that enters the  $IH^2$  process after being dried from 80% moisture. The data was divided into different sections. The first section includes product yields in which the two main products were the  $IH<sup>2</sup>$  renewable diesel and gasoline. The second main section is the raw materials which encompassed the dry biomass and total catalyst which includes the catalyst used for hydropyrolysis and hydroconversion. This catalyst is used for removing all oxygen. Other inputs in this section are the cooling water chemicals plus the boiler feed water chemicals (BFW). The third main section is the utilities section electricity used to run the  $IH<sup>2</sup>$  process and natural gas used for drying of the algae. The fourth section is the waste products section which has  $CO<sub>2</sub>$  in exhaust that is produced from the reformer. Lastly there is the co-product section which includes water produced from the  $IH^2$  processes, ammonia and ammonia sulfate, which are all mixed in specific ratios so as to produce fertilizers for sale. These co-products results in a GHG reduction credit for the  $\overrightarrow{IH}^2$  life cycle using a displacement

allocation. Input tables are similarly organized for other feedstock-specific  $IH^2$  inputs below. The simplified process flow diagram is shown in appendix H.

The inventory data from Table 3.6 was input to SimaPro, the LCA software tool used for this evaluation. This input data is shown in Table 3.6, organized by major life cycle stage. In the results section, GHG emissions will be reported for each of the major life cycle stages. Each of the inputs shown in Table 3.6 was multiplied by an energy allocation factor (EAF) which was calculated to be 1 so that the inventory would be apportioned to the main products (renewable diesel and gasoline) as well as the coproducts, steam exported from the  $I\text{H}^2$  process. The energy allocation factor was calculated using a methodology to be presented next. GHG emissions for the electricity used in the  $IH^2$  process were the US average grid (eGRID 2011) using an ecoprofile in the ecoinvent™ database in SimaPro. The eGRID emissions are from the site of the power plant only, and do not include upstream and transmission loss effects. In order to compensate for this, the eGRID emissions were multiplied by a factor of 1.1 twice; once for upstream processes (10% additional inventory) and a second time for transmission losses (10% loss assumed).

#### **Table 3.5**

**Data inputs for wood and forest residue raw material collection, transportation and yard processing based on 1 dry short ton biomass with an assumption of \$6 per gallon of diesel fuel.** 

<b>Life Cycle Stage</b>	<b>Items Used</b>	<b>Amounts 500</b>	<b>Amounts</b>
		dry tons/day	1000 dry
			tons/day
<b>Collection (Raw</b>	Diesel	1.047 gallons	1.197 gallons
material Inputs)			
	Lubricating oil	$0.013$ gallons	$0.017$ gallons
	Grease	0.038 gallons	0.048 gallons
	Hydraulic fluids	$0.014$ gallons	0.018 gallons
	Gasoline	$0.039$ gallons	$0.050$ gallons
<b>Transportation</b>	Diesel	$0.678$ gallons	0.914 gallons
	Lubricating oil	$0.014$ gallons	0.017 gallons
	Hydraulic fluids	0.014 gallons	0.018 gallons
	Tubes of grease	0.038 gallons	0.048 gallons
<b>Yard processing</b>	Diesel	$0.122$ gallons	$0.160$ gallons
	Lubricating oil	$0.016$ gallons	$0.016$ gallons
	Hydraulic oil	$0.016$ gallons	$0.016$ gallons
	Tubes of grease	0.043 gallons	0.043 gallons
(note: US average	Electricity	29.8 kWh	29.8 kWh
grid)			

#### **Table 3.6**

#### **Aquaflow Bionomic IH<sup>2</sup> inputs and outputs inventory for 80% moisture microalgae feedstock reduced to 20% moisture. Basis: 1 day operation of 2,000 MAF metric ton/day feedstock plant operation.**



# **3.2.2 Inputs for Bagasse IH<sup>2</sup> Biofuels Production**

Table 3.7 shows the  $IH^2$  facility inputs and outputs for the life cycle inventory of bagasse biofuels. The data was provided by Terry Marker (GTI) and was based on a 2,000 metric ton (MAF) of bagasse input/day plant with feedstock moisture of 45%. The accuracy of the input data was verified by carrying out a mass and energy balance as shown in appendix B. The data was divided into different sections, similar to those described in section 3.2.1. The factor of 2 appearing converts inputs to the basis of 2,000 MAF mt/day from the original set of data for a 1,000 mt/day facility.

The export steam was calculated in two different scenarios

- i) Char is burned to produce steam.
- ii) Char is a co-product and exported from the product system.

Both of these scenarios affect the energy allocation calculation as shown below in section 4.2.

The bagasse was dried from 45% moisture to 20% moisture to enhance size reduction and  $IH<sup>2</sup>$  conversion. The energy for drying was supplied by steam generated by the exothermic reactions occurring in the hydropyrolysis and hydroconversion reactions and was accounted for in the energy balance calculations which yielded the net steam exported (provided by GTI).

The input data from Table 3.7 was entered into SimaPro 7.2, the LCA software tool used for this evaluation. Each of the inputs shown in Table 3.7 was multiplied by an energy allocation factor (EAF) which was 0.897 in the scenario where char is burned and 0.724 in the scenario which char is considered as a co-product. The inventory is allocated to the main products ( $IH^2$  diesel and gasoline), and the co-products, ammonia and ammonium sulfate, provide an environmental impact credit in this analysis. The energy allocation factor was calculated using a methodology to be presented in section 4.2.

#### **Table 3.7**

**IH<sup>2</sup> inputs and outputs for the 45% moisture bagasse feedstock. Basis is 1 day operation of 2,000 moisture and ash free (MAF) metric ton/day plant operation.** 



## **3.2.3 Inputs for Corn Stover IH<sup>2</sup> Biofuels Production**

Table 3.8 shows the IH<sup>2</sup> facility inputs and outputs provided for the life cycle assessment. The data was obtained from Terry Marker (GTI) and Eric Tan (NREL) and was based on a 2,000 dry metric ton/day plant based on an assumption of 20% moisture content of the corn stover biomass feedstock. The accuracy of input data was verified by carrying out a mass and energy balance as shown in appendix C. The data was divided into different sections as shown previously.

The input data from Table 3.8 was entered to SimaPro 7.2, the LCA software tool used for this evaluation. Each of the inputs shown in Table 3.8 was multiplied by an energy allocation factor (EA factor) which was calculated to be 0.755 so that the inventory would be apportioned to the main products (renewable diesel and gasoline) as well as the co-products, steam exported from the  $IH^2$  process.

#### **Table 3.8**

#### **IH<sup>2</sup> inputs and outputs inventory for the 20% moisture corn stover feedstock. Basis is 1 day operation of 2,000 moisture and ash free (MAF) metric ton/day plant operation.**





**Forest resources IH<sup>2</sup> inputs and outputs inventory for the 30% moisture and 50% moisture feedstock. Basis: 1 day of operation of 2,000 dry metric ton/day facility.** 



## **3.2.4 Inputs for Forest Resources IH<sup>2</sup> Biofuels Production**

Table 3.9 shows the  $IH^2$  facility inputs and outputs provided for the life cycle assessment for the Johnson Timber Company's forest feedstock. The inventory data was obtained from Terry Marker (GTI) and Eric Tan (NREL) and was based on a 2,000 dry metric ton/day  $IH^2$  plant with feedstock dried to moisture of 10%. This data was based on an assumption of 30% and 50% feedstock moisture for two separate scenarios. The accuracy of the input data was verified by carrying out a mass and energy balance as shown in appendix D. This data was divided into different sections, similar to Table 3.3 in section 3.2.1.

The inventory data from Table 3.9 was input to SimaPro, the LCA software tool used for this evaluation. In the results section, GHG emissions will be reported for each of the major life cycle stages. Each of the inputs shown in Table 3.9 was multiplied by an energy allocation factor (EAF) so that the inventory would be apportioned to the main products (IH<sup>2</sup> diesel and gasoline) as well as the co-products, ammonia and ammonium

sulfate. The energy allocation factor was calculated using a methodology to be presented next. GHG emissions for the electricity used for the grinding and the  $\overrightarrow{IH}^2$  process were the US average grid. The original data was obtained from the table in appendix F.

# **4. Energy Allocation**

Energy allocation (EA) was applied in order to distribute the system environmental burdens among all products and co-products in the  $IH^2$  biofuel production chain. The EA method includes an energy balance utilizing material flows and lower heating values (LHV) for each co-product from the  $IH^2$  biofuel conversion stage. No co-products were generated in any other stage for all of the feedstocks considered in this study. The following sections describe the calculations made to determine energy allocation factors  $(EA$  factor) to be applied to allocate environmental impact to the main  $IH^2$  biofuel products. The EA factor was applied to all inputs in every life cycle stage to the  $IH<sup>2</sup>$ biofuels production system. Energy allocation is an energy balance around the  $IH<sup>2</sup>$ process where co-products are produced. We wish to know what fraction of total output of energy from the process is contained in  $IH^2$  biofuels. Energy can be carried out of the process in various forms;  $IH^2$  biofuels, steam, and char co-product. As a quality check on these energy balance calculations, we also attempted to balance the total input energy from the input biomass to the  $IH^2$  conversion process, with all output energy streams. Our attempts to do this from the data provided by GTI yielded energy balances that did not close perfectly, but the output energy was lower than the input energy by 5-20% for most feedstocks. Although this is not perfect data quality, such a result is consistent with energy losses from the process in the form of waste heat which was not quantified. In summary, we feel that the data quality was of sufficiently high quality to proceed with the final analyses.

The (EA) factor was obtained by using the equations below whereby the denominator represents the total energy out from all products and numerator is energy content of the  $\overrightarrow{IH}^2$  gasoline and IH<sup>2</sup> diesel.

$$
EA_{Biomass\,\,Feedstock} = \frac{Energy\,\,Out\,(\,\,gasoline+diesel)}{\,\,Total\,\,Energy\,\,out}
$$

# **4.1 Microalgae IH2 Biofuels**

When the individual inputs are included the above equation transforms into;

#### **20% Moisture Content Micro algae**

$$
EAF = \frac{\frac{gasoline}{\left(44\frac{MJ}{kg} * 4.48 \times 10^5 \frac{kg}{day}\right) + \left(44\frac{MJ}{kg} * 4.48 \times 10^5 \frac{kg}{day}\right)}}{\frac{\left(44\frac{MJ}{kg} * 4.48 \times 10^5 \frac{kg}{day}\right) + \left(44\frac{MJ}{kg} * 4.48 \times 10^5 \frac{kg}{day}\right)}{gasoline}} = 1
$$
# **4.2 Bagasse IH<sup>2</sup> Biofuels**

When the individual inputs are included the above equation transforms into;

#### **20% moisture content bagasse with char as a product**





#### **20% moisture content bagasse with char burned**



The lower heating values (LHV) of the fuels, steam, and char were obtained from existing databases in the MTU LCA group.

# **4.3 Corn Stover IH2 Biofuels**

When the individual inputs are included the above equation transforms into;

#### **20% moisture content corn stover with char burned**



# **4.4 Forest Resources IH2 Biofuels**

For the two different feedstock moisture scenarios, the energy allocation factor equations are as seen in the equations below. The Low Heating Value of the hydrogen was obtained from literature (Grohmann et al. 1984), while the LHV for the wood biomass was obtained from other literature.

### **30%moisture feedstock**

$$
EA\,factor = \frac{\overbrace{\left(44\frac{MJ}{kg} \times 3.2 \times 10^{5} \frac{kg}{day}\right)}^{gasoline} + \overbrace{\left(44\frac{MJ}{kg} \times 1.99 \times 10^{5} \frac{kg}{day}\right)}^{diesel}}^{\text{gasoline}}}{\underbrace{\left(44\frac{MJ}{kg} \times 3.2 \times 10^{5} \frac{kg}{day}\right)}_{gasoline} + \underbrace{\left(44\frac{MJ}{kg} \times 1.99 \times 10^{5} \frac{kg}{day}\right)}_{diesel} + \underbrace{\left(1.9845\frac{MJ}{kg\,steam} \times 2.92 \times 10^{5} \frac{kg}{day}\right)}_{steam} = 0.798}
$$

## **50% moisture feedstock**

$$
EA factor = \frac{\frac{gasoline}{(44\frac{MJ}{kg} \times 3.2 \times 10^5 \frac{kg}{day})} + \frac{diesel}{(44\frac{MJ}{kg} \times 3.2 \times 10^5 \frac{kg}{day})} + \frac{44\frac{MJ}{kg} \times 1.99 \times 10^5 \frac{kg}{day})}{(44\frac{MJ}{kg} \times 2.0 \times 10^5 \frac{kg}{day})} + \frac{44\frac{MJ}{kg} \times 2.0 \times 10^5 \frac{kg}{day})}{(1.9845\frac{MI}{kg} \times 1.45 \times 10^5 \frac{kg}{day})}} = 0.888
$$

# **5. Life Cycle Impact Assessment**

The inventory data were converted to greenhouse gas impacts using the IPCC GWP 100a method in SimaPro 7.2. This method converts emissions of greenhouse gases into equivalent emissions of  $CO<sub>2</sub>$  by employing global warming potentials (GWP). The GWP of CO<sub>2</sub> is 1, for CH<sub>4</sub> = 25, and for N<sub>2</sub>O is 298. Other greenhouse gases are also included in this analysis, including solvents and refrigerants that accompany ecoprofiles resident in SimaPro and called into the analysis with the material and energy inputs. In this study,  $CO<sub>2</sub>$  emissions from (or sequestration into) biogenic carbon were not counted in the GHG analysis, only fossil derived  $CO<sub>2</sub>$  emissions. The reason for this distinction between biogenic and fossil carbon is because when biomass grows,  $CO<sub>2</sub>$  is sequestered from the atmosphere into biomass, and upon conversion and combustion of biofuels, the  $CO<sub>2</sub>$  is returned to the atmosphere again in a closed cycle. We did not take a consequential view of the fate of biogenic carbon in the  $IH^2$  pathway, for example if CH<sub>4</sub> emissions would be a result of changes in biogenic carbon throughout the life cycle. As of this writing, field data was lacking to provide such data, and we believe that is will be a minor contributor to the life cycle. The inputs for corn stover production were acquired from the farms as a waste and this was done within the sustainable limit by only taking what we required and left some of the of the crops in the soil to degrade and replenish some of the lost nutrients and soil organic matter for the next crop.

# **5.1 Microalgae IH2 Biofuel**

The results from the SimaPro analysis were arrived at by dividing the 1-day impact results by the total energy content of the  $IH^2$  biofuels produced (39,424,000 MJ/day), or multiplying by the reciprocal which was 2.54E-8 of a day/MJ. This calculation is shown equations below.

$$
\frac{1}{IH^2\ gasoline + IH^2\ diesel}
$$
  

$$
\frac{1}{(19,712,000\ MJ/day) + (19,712,000\ MJ/day)} = 2.54E - 8\ day/MJ
$$

Doing this converted the GHG emissions from a 1 day basis to 1 MJ IH<sup>2</sup> biofuel basis.

# **5.2 Bagasse IH<sup>2</sup> Biofuels**

The results from the SimaPro analysis were arrived at by dividing the 1-day impact results by the total energy content of the  $IH^2$  biofuels produced (25,225,200 MJ/day), or multiplying by the reciprocal which was 3.96E-8 of a day/MJ. This calculation is shown equations below.

$$
\frac{1}{IH^2\ gasoline + IH^2\ diesel}
$$
  
25

$$
\frac{1}{(19,051,200 \text{ MJ/day}) + (6,174,000 \text{ MJ/day})} = 3.96E - 8 \text{ day/MJ}
$$

Doing this converted the GHG emissions from a 1 day basis to 1 MJ  $\text{IH}^2$  biofuel basis. A comparison of the GHG results for  $IH^2$  biofuels is compared to the life cycle GHG emission for petroleum gasoline, diesel, and aviation fuel.

# **5.3 Corn Stover IH2 Biofuels**

The results from the SimaPro analysis were arrived at by dividing the 1-day impact results by the total energy content of the  $IH^2$  biofuels produced (22,880,000 MJ/day), or multiplying by the reciprocal which was 4.37E-8 day/MJ. This calculation is shown equations below.

$$
\frac{1}{IH^2\ gasoline + IH^2\ diesel}
$$
  

$$
\frac{1}{(14,080,000\ M]/day) + (8,800,000\ M]/day)} = 4.37E - 8\ day/MJ
$$

# **5.4 Forest Feedstocks IH2 Biofuels**

The results from the SimaPro analysis were arrived at by dividing the 1-day impact results by the total energy content of the  $IH^2$  biofuels produced (22,880,000 MJ/day), or multiplying by the reciprocal which was 4.37E-8 day/MJ. This calculation is shown equations below.

$$
\frac{1}{IH^2\ gasoline + IH^2\ diesel}
$$
  

$$
\frac{1}{(14,080,000\ MJ/day) + (8,800,000\ MJ/day)} = 4.37E - 8\ day/MJ
$$

# **6. Life Cycle Assessment Results**

# **6.1 Microalgae Biomass and IH2 Biofuel Results**

### **6.1.1 Microalgae Biomass Production**

The results obtained from this analysis are grouped into four main sections: i. Algae Production Pump Shed ii. Algae Production New Harvest Units, iii. Algae Production Dewatering, and iv. Algae Transport. Figure 6.1 shows the GHG emissions per dry metric ton algae produced assuming 300 g algae/m<sup>3</sup> cell density. The Pump Shed stage emits the largest amount of emissions, followed by Algae Production Dewatering, Algae Transport, and Algae New Harvest Units. Table 6.1 shows the effects of primary energy type on the electricity impacts of producing algae. Coal electricity emits the largest amount of emissions, followed by US average grid and natural gas, with renewable electricity emitting the least.



**Figure 6.1 Greenhouse gas emissions per dry metric ton algae biomass (657 kg CO2 eq. / metric ton algae) assuming average US grid electricity.** 



**Effect of Electricity Type (Primary Energy) on GHG Emissions of Algae** 

## **6.1.2 Microalgae IH2 Biofuel Production and Use**

The inputs listed in Table 3.2 were entered into a project in SimaPro in order to determine the greenhouse gas emissions per MJ of  $IH<sup>2</sup>$  biofuels produced and used in vehicles. Figure 6.2 shows the total GHG emissions of .0619 kg CO<sub>2</sub> eq./MJ IH<sup>2</sup> biofuels, or 61.9 g  $CO<sub>2</sub>$  eq./MJ. To place these emissions into perspective, petroleum gasoline has life cycle GHG emissions of 91.2 g CO<sub>2</sub> eq./MJ. This  $\text{H}^2$  biofuel result was obtained assuming US average grid electricity used for algae feedstock production and also for electricity use during  $IH^2$  biofuel production (IH<sup>2</sup> processes in Figure 6.2). The largest contributor to emissions is algae feedstock production and transport to the  $IH^2$ facility, followed by  $IH^2$  processes for producing biofuels. Natural gas combusted for drying algae from 80% to 20% is the largest single cause of GHG emissions and electricity use for algae harvesting and dewatering is also a major cause for emissions. The emission credits from co-products ammonia and ammonium sulfate total about 20% of the net GHG emissions. The GHG results in Figure 6.2 include effects of biofuels combustion, but do not include transport of  $I\text{H}^2$  biofuels to blending locations for mixing into petroleum fuel stocks, nor from the blending location to filling stations. The latter step is considered negligible based on prior experience with biofuel life cycles, and therefore is omitted from this study.

Electricity type has a large impact on GHG emissions as shown in Table 6.1, and similarly has a large effect on  $\overline{H}^2$  biofuel emissions as shown in Table 6.2. When coal electricity is used, emissions are highest at 82.8 g  $CO<sub>2</sub>$  eq./MJ and are least when a renewable power source is used such as hydroelectric power; 37.9 g  $CO<sub>2</sub>$  eq./MJ. There is a very strong influence of electricity type on these GHG results. When mode of transportation from  $I H^2$  facility gate to blending location assuming 100 km distance is explored, there is very little difference between the transport modes.

**Effect of Electricity Type on IH2 Biofuel GHG Emissions** 



## **Table 6.3**

### **Effect of Transport Mode to Blending Location on IH<sup>2</sup> Biofuel GHG Emissions. Electricity Type is US Average Grid Power.**







# **6.1.3 Comparison of different scenarios for the transportation of IH2 fuel to blending stations.**

The different scenarios considered for the  $IH^2$  fuel transportation from gate to a blending station and other changes based on the different scenarios which will be discussed in this report are as follows:

- 1. Transportation of  $IH^2$  biofuel to a blending station using different modes of transport.
- 2. Transportation of IH<sup>2</sup> biofuel to a blending station using road transport over different distances.

### **6.1.4 Transportation Mode Scenario:**

Transportation of the  $IH^2$  biofuel to one selected blending station using different modes of transport. Tables 6.4 are the results obtained from the different transportation modes considering 20% feedstock moisture content. The different modes of transport were: a) Rail b) Road c) Pipeline

**Assumption:** The  $IH^2$  biofuel transportation distance from facility to filling station is 100 km.

#### **Discussion:**

It is assumed that the  $IH^2$  biofuel fuel shares similar properties as their respective fossil fuels used. The transportation of the  $IH^2$  biofuel using the pipeline mode has the least GHG emissions at  $61.9g$  CO<sub>2</sub> eq./ MJ of fuel. The emission from the rail transport was moderate at 8.99E-5g  $CO<sub>2</sub>$  eq. / MJ of fuel. The highest emission was from road transport, at 1.25g  $CO_2$  eq./ MJ of fuel, with a cumulative total emission of 62.2g  $CO_2$ eq./ MJ of IH<sup>2</sup> fuel. These transport numbers and total IH<sup>2</sup> biofuel GHG emissions can be seen in Table 6.4 below. The  $IH^2$  fuel transport step adds negligible amount of GHG emissions to the total GHG emissions, regardless of transport mode assuming 100 km distance.



**GHG emissions for the IH<sup>2</sup> process with 20% moisture Micro algae feedstock for different transport modes for IH2 product.**

# **6.1.5 Distance of Road Transport to IH2 Fuel Blending Facility Scenario:**

In this scenario an estimation of the effects of different transportation distances on GHG emissions, from the location of the  $IH^2$  biofuel production facility to different blending sites using road transport, will be made. We will use the same distances as in the Johnson Timber  $IH^2$  LCA report for this Micro algae analysis. The transportation distances to the various blending sites are shown in Table 6.5. The results obtained from the different transportation locations of the  $IH^2$  biofuel are shown below in Table 6.5.

#### **Table 6.5**

#### **Estimated distances for different blending locations**



## **6.1.6 Discussion of Micro Algae IH<sup>2</sup> Biofuel LCA Results**

From Table 6.6 below, the GHG emissions contribution from the  $IH<sup>2</sup>$  biofuel transport section varies with the  $IH^2$  biofuel road transportation distances for Micro algae feedstocks. This also directly affects the total GHG emissions achieved from the analysis. From Figure 6.2 above, it is clear that two major inputs dominate the GHG emissions; electricity for algae harvesting and natural gas for algae drying. The effects of uncertainty in these inputs can affect life cycle GHG emissions. For example, if electricity inputs for harvesting are increased or decreased by 50% to represent uncertainty in this input,  $IH^2$  biofuel GHG emissions increase or decrease to 77.9 and 45.9 g  $CO_2$  eq / MJ, respectively. Similarly for natural gas uncertainty, IH<sup>2</sup> biofuel GHG emissions increase or decrease to 80.5 and 43.3 g  $CO<sub>2</sub>$  eq / MJ, respectively.

Savings of GHG emissions of  $IH^2$  biofuel compared to petroleum fuels is shown in Figure 6.3. IH<sup>2</sup> biofuels in this comparison are produced using coal, US grid, and hydro power, and savings of GHG emissions compared to petroleum gasoline are 8%, 32%, and 58%, respectively. It is clear from these results that significant savings of emissions are only possible when renewable power is utilized for algae harvesting and dewatering. However, further reductions in GHG emissions is still possible if a renewable energy source could be found for the natural gas required for drying the algae biomass from 80% - 20%. Possible candidates could be landfill gas, anaerobic digester gas, and solar drying.

#### **Table 6.6**

**GHG emissions for production and transportation of IH<sup>2</sup> biofuel produced considering a 20% moisture Micro algae feedstock for Aquaflow Bionomic Company to different blending sites**

	<b>GHG</b>	<b>GHG</b>	<b>GHG</b>	<b>GHG</b>
	<b>Emissions</b>	<b>Emissions</b>	<b>Emissions</b>	<b>Emissions</b>
<b>Life Cycle Stages</b>	(g CO <sub>2</sub> )			
	eq./MJ of	eq./MJ of	eq./MJ of	eq./MJ of
	$IH2$ fuel)	$IH2$ fuel)	$IH2$ fuel)	$IH2$ fuel)
$IH2$ Road Transport Distance	147 miles	202 miles	277 miles	392 miles
$IH^2$ Feedstock and				
Transportation	36.7	36.7	36.7	36.7
$IH^2$ Process	25.1	25.1	25.1	25.1
$IH2$ biofuel Transportation	0.56	0.774	1.06	1.5
<b>Total GHG Emissions</b>	62.4	62.6	62.9	63.4





# **6.2 Bagasse Biomass and IH2 Biofuel Results**

### **6.2.1 Bagasse Biomass Production**

The main categories of the bagasse handling which are considered for the LCA analysis were i) bagasse transportation ii) bagasse loading and unloading, and iii) bagasse energy. Figure 6.4 below shows the greenhouse gas emissions per dry metric ton for loading, unloading, and transportation to a  $IH^2$  unit 100 km distance from the sugarcane milling factory. The total GHG emissions are 27.1 kg  $CO<sub>2</sub>$  eq. per dry metric ton bagasse. The largest contributor to this total is the transportation process. The bagasse transportation is equivalent to 24.40 kg  $CO<sub>2</sub>$  eq. per dry metric ton secondly is the loading and unloading of the bagasse which is very low at 2.65 kg  $CO<sub>2</sub>$  eq. per dry metric ton which is about 15% of the total emissions.



**Figure 6.4 Network diagram with magnitudes of GHG emissions from Bagasse**  handling to the IH<sup>2</sup> process (kg CO<sub>2</sub> eq./dry mt bagasse).

## **6.2.2 Bagasse IH<sup>2</sup> Biofuel Production and Use**

The total GHG emissions for this feedstock where the **char is burned** for steam production, is 2.6 g CO<sub>2</sub> eq /MJ of IH<sup>2</sup> fuel produced, as shown in the Figure 6.5. The  $IH<sup>2</sup>$  feedstock handling and transportation accounts for most of the emissions, which is 1.92 g CO<sub>2</sub> eq /MJ of  $IH^2$  fuel produced. The lowest emissions are from the IH<sup>2</sup> process which is a credit of -0.892 g  $CO_2$  eq /MJ of IH<sup>2</sup> fuel produced, due to the emissions credits from ammonia and ammonium sulfate co-products. These emission credits were obtained from ecoprofiles in the ecoinvent database in SimaPro 7.2. The IH<sup>2</sup> feedstock onsite preparation is 1.57 g eq  $CO<sub>2</sub>/MJ$  of IH<sup>2</sup> fuel produced.



### **GHG emissions for the IH<sup>2</sup> process with bagasse feedstock**

The total GHG emissions for bagasse feedstock for **char as a product scenario** is 2.1 g  $CO<sub>2</sub>$  eq /MJ of IH<sup>2</sup> fuel produced, as shown in the Table 6.7. These results are very similar to the char burned case except slightly lower because of the lower EA factor (.724).



Figure 6.5 Network diagram for bagasse IH<sup>2</sup> biofuels emission of GHGs (kg CO<sub>2</sub> eq./MJ IH<sup>2</sup> biofuels). Figure 6.5 Network diagram for bagasse IH<sup>2</sup> biofuels emission of GHGs (kg CO<sub>2</sub> eq./MJ IH<sup>2</sup> biofuels). Red lines show relative magnitude of greenhouse gas impacts while green lines show credits due to co-**Red lines show relative magnitude of greenhouse gas impacts while green lines show credits due to co**products of production. Line width corresponds to magnitude of impact or credit. **products of production. Line width corresponds to magnitude of impact or credit.**



### **GHG emissions for the IH<sup>2</sup> process bagasse showing effects of 100 km transport of IH<sup>2</sup> fuel to blending stations by different transport modes**

In Table 6.8 are results obtained from the different  $IH^2$  biofuel transportation modes for the char burned base case. The different modes of transport were: a) Rail b) Road c) Pipeline. The IH<sup>2</sup> biofuel transportation distance from facility to filling station is 100 km. For this short distance, there is little effect of  $I\text{H}^2$  biofuel transport to blending stations.

In this scenario an estimation was made of the effects of different transportation distances on GHG emissions from the location of the  $IH^2$  biofuel production facility to different blending sites using **road transport**. We will use the same distances as in the Johnson Timber  $IH^2$  LCA report for this bagasse analysis.

The transportation distances to the various blending sites are shown in Table 6.9. The results obtained from the different transportation locations of the  $IH^2$  biofuel are shown below in Table 6.10.

<b>Different blending locations</b>	<b>Distances</b>
Scenario 1	147 miles
Scenario 2	202 miles
Scenario 3	277 miles
Scenario 4	392 miles

**Distances to different blending sites being considered**

<b>Table 6.10</b>	
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**GHG emissions for production and transportation of IH<sup>2</sup> biofuel produced assuming a 20% moisture bagasse feedstock to different blending sites** 



From Table 6.10 below, the GHG emissions contribution from the  $IH<sup>2</sup>$  biofuel transport section varies with the  $IH^2$  biofuel transportation distances for bagasse feedstocks. There is not much effect of distance to blending facility, even for the longest regional distance of 392 miles, on the total GHG emissions for bagasse  $IH<sup>2</sup>$  biofuels.

## **6.2.3 Discussion of Bagasse IH<sup>2</sup> Biofuel LCA Results**

A comparison was conducted between the GHG emissions of  $IH<sup>2</sup>$  biofuels from bagasse biomass emissions to the emissions from convectional petroleum gasoline, diesel, and jet fuel shown in Figure 6.6 below. It is clear that two major inputs dominate the GHG emissions; diesel for bagasse  $IH^2$  feedstock transportation and diesel onsite preparation for bagasse. The effects of uncertainty in these inputs can affect life cycle GHG emissions. For example, if transportation inputs for harvesting are increased or decreased by 50% to represent uncertainty in this input,  $IH^2$  biofuel GHG emissions increase or decrease to 3.47 and 1.74 g  $CO<sub>2</sub>$  eq / MJ, respectively. Similarly for onsite

feedstock preparation uncertainty,  $IH^2$  biofuel GHG emissions increase or decrease to 3.39 and 1.82 g CO<sub>2</sub> eq / MJ, respectively. These emissions from the bagasse  $IH^2$ biofuels are relatively low compared to the data from National Energy Technology Laboratory (NETL 2008). Savings of GHG emissions compared to petroleum fuels is approximately 97%, easily qualifying these biofuels as adavanced biofuels according to the Renewable Fuels Standard (50% reduction required).



**97% 98% GHG Savings Compared to Petroleum**

Figure 6.6 Results of IH<sup>2</sup> fuel for bagasse feedstock ghg emisisons results savings compared to petroleum fuels (no transport step to blending was included here**negligible effect).** 

# **6.3 Corn Stover Biomass and IH2 Biofuel Results**

#### **6.3.1 Corn Stover Biomass Production**

The main categories of the corn stover production system which are considered for the LCA analysis were i. fertilizer replacement, ii. corn stover collection, iii. corn stover transportation, and iv. corn stover loading and loading. Figure 6.7 shows the greenhouse gas emissions per dry metric ton of fertilizer replacement, collection, loading, unloading and transported to a  $\text{IH}^2$  unit 48 km (30 mi.) distant from the corn stover fields. The total GHG emissions are 66.8 kg  $CO<sub>2</sub>$  eq. per dry metric ton corn stover biomass. The largest contributor to this total is the fertilizer replacement, followed by collection, transport, and loading/unloading.



**Figure 6.7 Network diagram with GHG emissions from Corn Stover collection,**  loading, transport, and fertilizer replacement ( $kg CO<sub>2</sub>$  eq./dry mt stover).

## **6.3.2 Corn Stover IH<sup>2</sup> Biofuel Production and Use**

The total GHG emissions for corn stover IH<sup>2</sup> biofuel where the **char is burned** for steam production is shown in a network diagram in Figure 6.8. The largest emission is from corn stover production, followed by size reduction, and with a credit for co-products ammonia and ammonium sulfate. Several  $I H<sup>2</sup>$  biofuel transportation scenarios were studied assuming 100 km distance to locations of blending into petroleum fuel stocks; a) rail, b) road, and c) pipeline. Table 6.11 shows the results from these scenarios. Road transport adds about 5% to these base case emissions, but rail and pipeline transport contribute negligibly to the total emissions.

In another scenario an estimation was made of the effects of different transportation distances on GHG emissions from the location of the  $IH<sup>2</sup>$  biofuel production facility to different blending sites using **road transport**. We will use the same distances as in the Johnson Timber  $\overline{IH}^2$  LCA report for this bagasse analysis. The transportation distances to the various blending sites are shown in Table 6.12. The results obtained from the different transportation locations of the  $IH^2$  biofuel are shown below in Table 6.12. As in the bagasse case, there is not much effect of distance to blending facility, even for the longest regional distance of 392 miles, on the total GHG emissions for corn stover  $IH<sup>2</sup>$ biofuels.



**Figure 6.8** Network diagram for corn stover IH<sup>2</sup> biofuels emission of GHGs (kg CO<sub>2</sub> eq./MJ IH<sup>2</sup> biofuels). Red lines show relative magnitude of greenhouse gas impacts while green lines show credits due to co-products of **lines show relative magnitude of greenhouse gas impacts while green lines show credits due to co-products of**  Figure 6.8 Network diagram for corn stover IH<sup>2</sup> biofuels emission of GHGs (kg CO<sub>2</sub> eq./MJ IH<sup>2</sup> biofuels). Red production. Line width corresponds to magnitude of impact or credit **production. Line width corresponds to magnitude of impact or credit** 



**GHG emissions for the IH<sup>2</sup> process with 20% moisture feedstock corn stover** 

### **Table 6.12**

**GHG emissions for production and transportation of IH<sup>2</sup> biofuel produced considering a 20% moisture corn stover feedstock to different blending sites** 



## **6.3.3 Discussion of Corn Stover IH2 Biofuel LCA Results**

A comparison was conducted between the GHG emissions of  $IH<sup>2</sup>$  biofuels from corn stover biomass emissions to the emissions from convectional petroleum gasoline, diesel, and jet fuel shown in Figure 6.10. It is clear that two major inputs dominate the GHG emissions; diesel for cornstover  $IH^2$  feedstock collection and Fertilizer for replenishing the lost nutrients for the next crop to be grown. The effects of uncertainty in these inputs can affect life cycle GHG emissions. For example, if onsite preparation inputs are increased or decreased by 50% to represent uncertainty in this input,  $IH^2$  biofuel GHG emissions increase or decrease to 7.33 and 5.87 g  $CO<sub>2</sub>$  eq / MJ, respectively. Similarly for fertilizer replenishment increasing or decreasing by  $50\%$ , IH<sup>2</sup> biofuel GHG emissions increase or decrease to 9.67 and 3.53g  $CO<sub>2</sub>$  eq / MJ, respectively. These emissions from the corn stover  $IH^2$  biofuels are relatively low compared to the data from National Energy Technology Laboratory (NETL 2008). Savings of GHG emissions compared to petroleum fuels is approximately 93%, easily qualifying these biofuels as adavanced biofuels according to the Renewable Fuels Standard (50% reduction required).



**Figure 6.10 Results of IH<sup>2</sup> fuel for corn stover feedstock GHG emisisons results savings compared to petroleum fuels (no transport step to blending was included here-negligible effect).** 

# **6.4 Forest Resources Biomass and IH<sup>2</sup> Biofuel Results**

### **6.4.1 Forest Resources Biomass Production**

The results obtained from this analysis are grouped into three main sections: collection, transportation and yard preprocessing. Figure 6.11 shows a network diagram of the GHG impacts of these three sections on the basis of 1 dry metric ton assuming a 1,000 dry metric ton / day facility. The largest source if GHG emission is electricity consumed for size reduction of the biomass. Diesel fuel for biomass collection is the next largest, followed by diesel fuel for transportation.

Two sets of results were obtained, one for the 500 and 1,000 dry metric ton/day plants assuming diesel fuel costs of \$3/gallon, and another assuming \$6/gallon. These results are shown in Tables 6.13 and 6.14. The general trends are that emissions increase for the larger feedstock supply and for lower fuel prices. The reasons for these trends are that larger distances are needed for transport for the larger supply need, and for higher fuel prices, this favors collection of higher cost resources closer to the facility. These economic tradeoffs are possible with the forest procurement model provided by Mr. Gephardt, and the environmental tradeoffs are provided by the LCA.



**Figure 6.11 Network diagram with GHG emissions from Forest Feedstock collection, transport, and preprocessing (kg CO2 eq./dry mt forest resources) assuming a 1,000 dry metric ton/day and \$3/gallon diesel**  Figure 6.11 Network diagram with GHG emissions from Forest Feedstock collection, transport, and preprocessing (kg CO<sub>2</sub> eq./dry mt forest resources) assuming a 1,000 dry metric ton/day and \$3/gallon diesel **fuel.** 

**Greenhouse Gas Emissions per dry metric ton/day of wood and forest residues collected, transported, and processed on-site. Impacts of all greenhouse gases were**  converted to  $CO<sub>2</sub>$  equivalents using Global Warming Potentials (GWP). Plant sizes **of 500 and 1000 dry metric ton/day input feedstock considering electrical energy from US average grid as the yard processing energy source and assumption of \$3 per gallon of diesel fuel used.** 



**Greenhouse Gas Emissions per dry metric ton/day of wood and forest residues collected, transported, and processed on-site. Impacts of all greenhouse gases were**  converted to  $CO<sub>2</sub>$  equivalents using Global Warming Potentials (GWP). Plant size **of 500 and 1000 dry metric ton/day input feedstock considering electrical energy from US average grid as the yard processing energy source and assumption of \$6 per gallon of diesel fuel used.** 



## **6.4.2 Forest Resources IH2 Biofuel Production and Use**

A network diagram showing contributions to GHG emissions of  $IH<sup>2</sup>$  biofuels produced from 30% moisture content forest biomass is displayed in Figure 6.12. The largest emissions are from feedstock collection, transportation and size reduction  $(4.14 \text{ g } CO<sub>2</sub>)$ 

eq/MJ). Impacts from  $IH^2$  conversion process are very small, and an environmental credit is realized from co-products produced. Net GHG emissions are  $3.25 \text{ g } CO_2 \text{ eq/MJ}$ .

When 50% moisture content forest feedstocks are input to the  $IH<sup>2</sup>$  facility, GHG emissions are slightly higher as shown in Table 6.15. Slightly larger emissions are a result of a higher EAF applied in this case because a smaller amount of co-product steam is produced compared to the 30% moisture content case.

#### **Table 6.15**



**GHG emissions for the IH<sup>2</sup> process with 50% moisture feedstock content.** 

Transportation scenarios to deliver  $IH^2$  biofuel to a blending station located 100 km away using different modes of transport was studied. Table 6.16 contains these results. Road transportation has the highest impact, rail intermediate, and pipeline is the lowest. The effect of biofuel transport to blending locations is minimal.

More transport scenarios were studied by varying distance to blending locations assuming road transport. These distances were obtained by considering several blending facility locations in the Upper Midwest in the region surrounding Park Falls, WI, as shown in Table 6.17. GHG emissions for these transport scenarios are presented in Table 6.18. Even for the longest distance, additional emissions are only slightly larger than 1 g  $CO<sub>2</sub>$  eq/MJ.



**Figure 6.12 Network diagram for forest resources IH2 biofuels emission of GHGs (kg CO2 eq./MJ IH2 biofuels).**  Red lines show relative magnitude of greenhouse gas impacts while green lines show credits due to co-products **Red lines show relative magnitude of greenhouse gas impacts while green lines show credits due to co-products**  Figure 6.12 Network diagram for forest resources IH<sup>2</sup> biofuels emission of GHGs (kg CO<sub>2</sub> eq./MJ IH<sup>2</sup> biofuels). of production. Line width corresponds to magnitude of impact or credit. **of production. Line width corresponds to magnitude of impact or credit.** 



### **GHG emissions for the IH<sup>2</sup> process with 30% moisture forest resources assuming 100 km transport of IH<sup>2</sup> biofuel by different modes**

#### **Table 6.17**

### **Estimated distances for different blending locations**





**GHG emissions for production and transportation of IH<sup>2</sup> biofuel produced considering a 30% moisture forest residue feedstock for Johnson Timber Incorporated to different blending sites** 

## **6.4.3 Discussion of Forest Resources IH<sup>2</sup> Biofuel LCA Results**

A comparison was conducted between the GHG emissions of  $IH^2$  biofuels from forest biomass emissions to the emissions from convectional petroleum gasoline, diesel, and jet fuel shown in Figure 6.13. It is clear that two major inputs dominate the GHG emissions; diesel for Collection and transportation of  $IH<sup>2</sup>$  feedstock and electricity used for yard processing. The effects of uncertainty in these inputs can affect life cycle GHG emissions. For example, if the sum of the collection and transporation inputs are increased or decreased by 50% to represent uncertainty in this input,  $IH^2$  biofuel GHG emissions increase or decrease to 3.9 and 2.6 g  $CO<sub>2</sub>$  eq / MJ, respectively. Similarly for increase or decrease of electricity used for yard processing by  $50\%$ ,  $\text{IH}^2$  biofuel GHG emissions increase or decrease to 4.18 and 2.33 g  $CO<sub>2</sub>$  eq / MJ, respectively. These emissions from the forest resource  $IH^2$  biofuels are relatively low compared to the data from National Energy Technology Laboratory (NETL 2008). Savings of GHG emissions compared to petroleum fuels are approximately 96% for both the 30% and 50% moisture content biomass-based fuels, easily qualifying these biofuels as adavanced biofuels according to the Renewable Fuels Standard (50% reduction required).



Figure 6.13 Results of IH<sup>2</sup> fuel for forest feedstock GHG emisisons savings **compared to petroleum fuels (no transport step to blending was included herenegligible effect).**

## **7. Conclusions and Recommendations**

The uncertainty analysis was carried out by analyzing the highest impact inputs from LCA analysis results for each feedstock separately. The uncertainty results show a significant effect on the algae feedstock by making the greenhouse gas (GHG) emissions not to meet the 50% required potential greenhouse gas (GHG) savings required. Where's the uncertainty results for the other feedstocks does not make the feedstock not to meet the required standards.

The purpose of this report was to evaluate the cradle-to-grave life cycle assessment (LCA) of the Gas Technology Institute (GTI) Integrated Hydropyrolysis and Hydroconversion  $(IH^2)$  production chain, including the production of input feedstocks and use of output  $IH^2$  biofuels. This report contains a preliminary LCA based on input data for the production and delivery of biomass feedstocks to a future  $IH^2$  biofuel facility, and also based on inputs for the  $IH^2$  process provided by GTI. Alternative bio-based transportation fuels, such as the  $IH^2$  biofuels, have the potential to decrease climate change emissions from vehicular transportation. The goal is to compare environmental impacts of  $IH^2$  biofuels to equivalent fossil fuels in order to determine savings of greenhouse gas (GHG) emissions, but along this path, intermediate results for each biomass feedstock were generated and compared to each other. The functional unit for biomass feedstocks and  $\overline{IH}^2$  biofuels was 1 dry metric and 1 MJ of energy, respectively.

The main conclusion from this study is that GHG emissions for production and use of  $IH<sup>2</sup>$ biofuels from a variety of feedstocks (microalgae, cane bagasse, corn stover, forest resources) are very small compared to comparable petroleum fuels, with the possible exception of fuels derived from microalgae. Savings of GHG emissions per MJ of transportation fuels between 93-98% are typical of  $I\overline{H}^2$  biofuels produced from most of the studied biomass species (cane bagasse, corn stover, and forest resources). Explorations of  $IH^2$  biofuel transport modes (truck, rail, pipeline) and transport distances had very little effect on overall system GHG emissions. Microalgae produced using renewable electricity for collection and dewatering helped lower GHG emissions and increase savings above 50% compared to petroleum fuels, but the large energy burden of drying the high moisture microalgae feedstock (80% moisture) continues to be a challenge to approach the savings for bagasse, stover, and forest resources  $IH<sup>2</sup>$  biofuels.

In addition to these differences in GHG emissions for  $IH^2$  biofuels from several biomass feedstocks, there are also differences in biofuel production yields. Table ES2 shows yields of  $IH^2$  Biofuels from microalgae, cane bagasse, corn stover, and forest feedstocks. Microalgae  $IH^2$  biofuels exhibit the highest yields, nearly double the productivity of the other biomass feedstocks. Composition of biomass is likely the reason for these large differences in yields. For example, many species of microalgae contain significant oil, which contains fewer oxygen atoms and more hydrogen atoms per molecule. In such cases, a higher percentage of the starting biomass is expected to exit the process as biofuel as opposed to  $CO<sub>2</sub>$ ,  $H<sub>2</sub>O$  and other minor co-products. The yields in Table 7.1 also impact area productivity, that is, the quantity of biofuel produced per unit area of

surface of land or water per year. Area productivity is also affected by biomass productivity per unit surface area per year. Combining both of these productivities will result in a key indicator of overall biofuel production efficiency.

The results in this study represent a limited life cycle assessment that touched on one indicator of sustainability, greenhouse gas emissions and savings of those emissions compared to petroleum fuels. It is highly recommended to revisit this LCA when  $IH<sup>2</sup>$ conversion data is obtained on pilot or commercial scales. One topic of future interest might be LCAs of mixtures of these feedstocks for  $IH^2$  biofuel production; for example mixtures of microalgae and forest residue resources. Results from such future studies can help refine  $IH^2$  biofuel system impacts leading to more efficient production of this promising biofuel. It is also recommended to carry out scenario analysis considering the corn stover and bagasse as co-products not waste products. These future studies should also include other sustainability indicators for which little is known from this new transportation production system, including land use change emissions, water quantity and quality, emissions of other air pollutants, worker safety, community impacts from biomass transport, and employment. These expanded studies are particularly important when attempting to understand impacts of large-scale dissemination and implementation of this new renewable transportation fuels technology.







# **7.1.1 Discussion of comparison results of IH<sup>2</sup> base cases to the biofuel values for EPA 2010 regulation of fuels and fuel additives.**

Table 7.2 below shows a comparison between results for  $IH^2$  base cases biofuel values obtained from the energy allocation method to the EPA 2010 regulation of fuels and fuel additive values obtained using the displacement allocation method. This comparison is relevant because both the  $IH^2$  biofuels and the RFS2 biofuels listed are intended to displace petroleum gasoline and diesel in the market. With the exception of algae  $IH<sup>2</sup>$ biofuels, the results from  $IH^2$  process were much lower than the EPA 2010 results. The EPA values obtained using the displacement allocation method which generally gives

more favorable emission results as compared to the energy allocation method that was used for the IH<sup>2</sup> biofuels. The main reasons for the more favorable result for the IH<sup>2</sup> biofuels is because the feedstocks for sugar cane ethanol, corn ethanol, and soya bean biodiesel incur emissions from land use change, which is not the case for the  $IH<sup>2</sup>$ feedstocks, as explained previously in this thesis. A second reason for the more favorable result for  $IH^2$  biofuels is due to the processing differences. IH<sup>2</sup> processing employed process integration where hydropyrolysis and hydroconversion are integrated, allowing for more efficient use of energy.

#### **Table 7.2**





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### **Appendix**

7.2.1 Energy, mass balance and EAF calculations **7.2.1 Energy, mass balance and EAF calculations** 







## **Figure 7.2 Bagasse with char burnt energy, mass balance and EAF calculation**  Figure 7.2 Bagasse with char burnt energy, mass balance and EAF calculation















Figure 7.6 Forest Residue 50% moisture feedstock energy, mass balance and EAF calculation **Figure 7.6 Forest Residue 50% moisture feedstock energy, mass balance and EAF calculation** 





<sup>65</sup>

7.2.2 Original input data used for Life Cycle Analysis **7.2.2 Original input data used for Life Cycle Analysis** 

### Table 7.3 **Table 7.3**

# Original Johnson Timber feedstock Input table 2000 dry metric ton plant size **Original Johnson Timber feedstock Input table 2000 dry metric ton plant size**





**Original Johnson Timber IH2 Input tables table B: – Overall YE 2011-3Yields**

#### **Original Johnson Timber IH2 Input tables Feedstock Properties Typical Wood Yield 2011-3**



#### **7.2.3 Detailed General flow diagram of GTI IH2 process**



**Figure 7.8 More detailed Simplified flow diagram of GTI IH<sup>2</sup> process. (Diagram reference from extended abstract 2009 AIChE by Terry Marker, Larry Felix and Martin Linck from GTI)** 

**7.2.4 IH2 process diagram for algae conversion to biofuel** 

#### **Integrated Hydropyrolysis and Hydroconversion Process**



Figure 7.9 Simplified flow diagram of GTI IH<sup>2</sup> process. (Diagram reference from **extended abstract 2009 AIChE by Terry Marker, Larry Felix and Martin Linck from GTI)** 

**7.2.5 IH2 Process Flow diagram for Micro Algae feedstock production** 



**Figure 7.10 The process flow diagram for micro algae production from Aquaflow Bionomic Corporation. The shaded boxes represent the main stages of production.**

#### **Algae IH2 Input tables Overall Yields from 20% Moisture Algae Table D: YE 2011- 7**



 $*$  This  $CO<sub>2</sub>$  contains some Oxygen from air which is added when part of the reformer feed gas is burned in the reformer furnace.

Note: An additional 4189 t/d of moisture would be removed during drying from 80% water to 20% water



#### **Original Algae IH2 Input tables Typical Aquaflow Algae Table E: YE 2011-7**

Wastewater Algae Based on Aquaflow Algae mechanically died to 80% moisture; natural gas dried to 20% moisture

#### **7.2.7 Bagasse IH2 Input tables Overall Yields**

#### **Table 7.8**

#### **Original Bagasse IH2 Overall Yields from 20% Moisture Bagasse Table F: YE 2011- 8**



Product Yields

 $*$  This  $CO<sub>2</sub>$  contains some Oxygen from air which is added when part of the reformer feed gas is burned in the reformer furnace

Note: An additional 631 t/d of moisture would be removed during drying from 45% water to 20% water

#### **Estimated Overall Utilities starting with 45% moisture bagasse – char product made in table 3.7**



#### **Table 7.10**

#### **Estimated Overall Utilities staring with 45% moisture bagasse– char burned in hog boiler in table 3.7**

